

Influence of mechanical mixing rates on sludge characteristics and membrane fouling in MBRs

S. Jamal Khan and C. Visvanathan

Environmental Engineering and Management Program, School of Environment, Resources and Development,

Asian Institute of Technology, P.O. Box 4, Klong Luang, Pathumthani 12120, Thailand

Tel: (66-2)-524-5640; Fax: (66-2)-524-5625; E-mail: visu@ait.ac.th

V. Jegatheesan

School of Engineering, James Cook University, Townsville, Queensland 4811, Australia

R. Ben Aim

UMR5504, UMR792 Ingénierie des Systèmes Biologiques et des Procédés, CNRS, INRA, INSA,

F-31400, Toulouse, France

Abstract: The influence of shear intensity (G) induced by mechanical mixing on activated sludge characteristics as well as membrane fouling propensity in membrane bioreactors (MBRs) was investigated. Four MBRs were operated at different mechanical mixing conditions. The control reactor (MBR_0) was operated with aeration only supplemented by mechanical stirring at 150, 300 and 450 rpm in MBR_{150} , MBR_{300} and MBR_{450} , respectively. It was found that the MBR_{300} demonstrated minimum rate of membrane fouling. The fouling potential of the MBR_{300} mixed liquor was lowest characterized by the specific cake resistance and the normalized capillary suction time (CST_N). Moreover, it was found that the mean particle size reduced with increase in the shear intensity. These results reveal that membrane fouling can be significantly mitigated by appropriate shear stress on membrane fibers induced by mechanical mixing condition.

Keywords: Membrane fouling; Shear intensity; Particle size distribution; Extracellular polymeric substances (EPS); Specific cake resistance

INTRODUCTION

Membrane bioreactor (MBR) offers several advantages compared with conventional wastewater treatment processes including high biodegradation efficiency, excellent quality of effluent, smaller sludge production and compactness (1). However, the wide spread application of the MBR process is constrained by membrane fouling and it is considered as the most serious problem affecting system performance. Fouling results in permeate flux decline due to the interaction of membrane and activated sludge leading to frequent membrane cleaning and necessary membrane replacement. Fouling of membranes in MBRs is determined by three factors, namely: the nature of the feed to the membrane, the membrane properties and the hydrodynamic condition experienced by the membrane (2).

So far, several techniques for fouling control have been investigated including sub-critical flux operation, intermittent suction and backwashing (3). Membrane scouring with bubbling has been an effective hydrodynamic technique for fouling reduction in submerged MBRs. In recent researches, MBRs were operated with diffuser at the base to maintain aerobic condition and additional diffuser for air scouring of the membrane module (4-5). Lee et al. (6) found that membrane fouling in terms of rate of trans-membrane pressure (TMP) rise was dependent on air flow rate and TMP rose up more slowly with the increase in air flow rate. However, the high aeration intensity necessary to provide effective bubbling also leads to changes in the growth rate, the F/M ratio and the microbial community of sludge (7). Moreover, hollow fiber (HF) modules possessing high membrane surface area to footprint ratio are prone to excessive fouling due to the poor hydrodynamic conditions within the fiber bundles. It was found that the axial velocities induced by bubbling within HF module could be up to 10 times lower than the one outside the fiber bundle and the surrounded (center) fibers performed poorly compared to the outer fibers (8). High bundle packing density causes heterogeneous scouring

with bubbling and low shear intensity on surrounded fibers leads to solids accumulation between the fibers in the bundle. The present study was aimed at investigating the influence of mechanical mixing on HF membrane fouling propensity and on physical properties of the activated sludge. The membrane fouling behaviors and the sludge suspension characteristics from the MBRs at different mixing conditions were compared to determine the optimum mixing rate.

MATERIALS AND METHODS

Experimental setup and Operation

HF membrane modules (Mitsubishi Rayon) were submerged in bioreactors with 10 L working volume. The HF membranes were made of polyethylene having a pore size of 0.1 μm and an effective membrane filtration area of 0.42 m^2 . Synthetic wastewater simulating municipal wastewater was used as a substrate in the biological process with COD:N:P ratio of 100:10:2 and an organic loading rate (OLR) of 2.4 $\text{kg}/\text{m}^3\cdot\text{d}$. The composition of synthetic wastewater included dextrose (516 mg/L), soya protein (250 mg/L), NH_4Cl (229 mg/L), KH_2PO_4 (70 mg/L), CaCl_2 (10 mg/L), MgSO_4 (10 mg/L) and FeCl_3 (3 mg/L). pH in the bioreactors was maintained between 7.0 and 7.5 using NaHCO_3 (750 mg/L). Domestic activated sludge was acclimatized over a period of two months before seeding of the MBRs.

A constant air supply by filtered compressed air through air diffusers was maintained at a flow rate of 5 L/min in all the MBRs. Based on cross-sectional area of bioreactor, the air flow rate was equivalent to an aeration intensity of 10.6 $\text{m}^3/\text{m}^2\cdot\text{h}$. Dissolved oxygen (DO) in the MBRs was maintained between 2 and 5 mg/L. The varying condition among the four MBRs was the mechanical mixing with no stirring in control reactor (MBR_0) followed by stirring at

150, 300 and 450 rpm in MBR₁₅₀, MBR₃₀₀ and MBR₄₅₀, respectively. The membrane filtration was operated in a cyclic mode (10 min on, 2 min off) at a constant flux to maintain a hydraulic retention time (HRT) of 8 h. For submerged MBRs, intermittent suction is an effective approach for suppression of fouling (9). The permeate suction pressure was recorded using digital manometers connected to the suction line of the membrane modules.

Analytical Methods

Mixed liquor suspended solids (MLSS), volatile mixed liquor suspended solids (MLVSS), chemical oxygen demand (COD) and capillary suction time (CST) were determined according to APHA (10). Particle size distribution (PSD) of sludge samples was determined by light scattering technique using Mastersizer S (Malvern, UK). The particle size range was measured between 0.05 and 750 μm with an instrument accuracy of $\pm 1\%$.

EPS Analysis

Mixed liquor samples of 50 mL from the four MBRs were taken and cooled immediately at 4°C to minimize microbial activity. Soluble EPS was obtained by centrifugation of the mixed liquor at 4000 g for 20 min followed by high speed centrifugation at 20,000 g for 20 min and separation of the supernatant (2). Bound EPS was extracted from the mixed liquor using cation exchange resin (CER) extraction method (11). The CER (DOWEX HCR-S/S, Dow Chemical Company, USA) used was in Na⁺ form with bead size distribution range between 16-50 mesh. The centrifuged sludge was re-suspended in phosphate buffer solution and the CER (70 g CER/g VSS) was added and mixed at 600 rpm for 1 h. Then the mixture was centrifuged twice at 4000 g for 10 and 20 min, respectively, to obtain the supernatant as bound EPS. Carbohydrate and protein fractions of the soluble and bound EPS were measured by the colorimetric methods of Dubois et al. (12) and Lowry et al. (13), respectively.

D-Glucose and Bovine serum albumin (BSA) were used as carbohydrate and protein standards, respectively.

Determination of membrane fouling

The extent of membrane fouling in the MBRs was monitored in terms of rise in TMP with operational time. In this regard, flux and TMP were recorded on regular basis. The membrane fouling rates (dTMP/dt) were determined from the TMP profiles. The operation was stopped when TMP reached 30 kPa and chemical cleaning procedure was carried out. Prior to the chemical cleaning, the membrane unit was physically washed with tap water to remove visible cake layer from the membrane fibers. Then the membrane unit was immersed for 8 h in a solution constituting NaOCl (effective chlorine concentration of 3,000 mg/L) and 4 % aqueous NaOH. Following immersion period, the membrane unit was thoroughly rinsed with water to remove the chemical. This chemical cleaning protocol suggested by the membrane supplier (Mitsubishi Rayon) was able to recover intrinsic permeability 90-95 %.

Besides on-line data, batch filtration tests were performed to determine the specific cake resistance (α) of the sludge samples. The test was conducted in a 400 mL unstirred filtration cell (Model 8400, Amicon, USA) using a 0.22- μ m flat-sheet cellulose membrane filter (GVWP 09050, Millipore, USA). The cell was filled with 200 mL of mixed liquor sample and a constant pressure of 30 kPa was applied by pressurized nitrogen from a gas cylinder. The filtrate was continuously recorded by an electronic balance connected to a notebook using WINWEDGE software. The specific cake resistance, α (m/kg) was calculated (9) by

$$\alpha = \frac{2000 A^2 \Delta P t / V}{\mu C} \quad (1)$$

where ΔP is the applied pressure (kPa), A is the filtration area (0.00418 m^2), C is the MLSS concentration (kg/m^3), μ is the viscosity of permeate (N-s/m^2) and $[(t/V)/V]$ (s/m^6) is the slope of the straight portion of the curve that is obtained by plotting the time of filtration to volume of filtrate (t/V) versus the filtrate volume (V).

RESULTS AND DISCUSSION

The shear intensity in the MBRs was quantified by the mean velocity gradient (G) using expressions (14) presented in Table 1 and values reported in Table 2. According to Table 2, the pneumatic mixing due to air supply remained constant while the mechanical mixing due to stirring varied resulting in G variation among the MBRs.

The MBRs were run in a steady-state condition over a period of 120 days and the values of all the parameters were averaged along with standard deviation. The MLSS concentration was maintained between 6-8 g/L with MLVSS/MLSS ratio of approximately 90 % in the MBRs at sludge retention time (SRT) of 40 d. The COD removal efficiency of the MBRs was above 95% representative of effective biodegradation and physical separation by the HF membranes.

Filtration behavior in the MBRs

The variable mixing intensities in the MBRs resulted in membrane filtration behaviors as shown in Figure 1. It shows the rise in TMP for each MBR over a typical filtration period. It was observed that the membrane in MBR₀ fouled rapidly followed by the one in MBR₁₅₀. However, membrane filtration in the MBR₃₀₀ and MBR₄₅₀ could be achieved up to five times the filtration period of MBR₀. Taking into consideration the relatively similar biomass concentrations among the MBRs, MBR₃₀₀ demonstrated minimum fouling in terms of the filtration duration. Moreover, filtration duration could not be further increased in MBR₄₅₀.

with a higher G as compared to the one in MBR_{300} . In Figure 1, all the TMP profiles exhibited two-stage process. Initially, linear gradual TMP rise was observed followed by sudden increase in the rate of TMP rise leading to need for membrane chemical cleaning.

The two-stage process in membrane fouling behavior has been extensively investigated and explained by Cho and Fane (15) and later by Zhang et al. (2). Prior to these two-filtration steps, a conditioning period has been observed due to the initial adsorption of colloids and organics mostly before cake layer formation initiates under sub-critical flux operation (2). After membrane conditioning of MBRs, the bioflocs initiate cake formation on membrane fibers and between fibers in low shear stress regions of the HF bundle during the first stage. Over time, the cake deposition worsens leading to depletion of effective pores resulting in TMP rise. At the end of this stage, exponential TMP increase is observed when the effective pores become critical and permeate productivity redistributes to the less fouled pores, for which local flux exceeds the critical flux (16).

There are two significant parameters during the first stage: the critical time (t_{crit}) over which the first stage is maintained and the fouling rate ($dTMP/dt$) during this stage (17). Figure 1 shows that the first stage of fouling was maintained until TMP reached 7 kPa in the four MBRs operation. The TMP profile data reveals that the t_{crit} was observed at approximately 30, 80, 130 and 140 h during operation of the MBR_0 , MBR_{150} , MBR_{300} and MBR_{450} , respectively. The longer duration maintained in the MBR_{300} and MBR_{450} for the first fouling stage can be mainly attributed to the high shear intensity on membrane fibers induced by mechanical stirring.

Membrane fouling rates

Based on membrane filtration performance in the MBRs (Figure 1), the membrane fouling rates ($dTMP/dt$) during the first and second fouling stages were determined. The first stage ranged from the start-up TMP of 3 kPa to 7 kPa and the second stage ranged from 10 kPa to the terminating TMP of 30 kPa. The fouling rates were determined by the slope of the linear curve from the TMP versus time plot as shown in Figure 2.

The first stage fouling rates are representative of pore blocking, biopolymer deposition, biofilm attachment and growth, all contributing to steady TMP rise (2). Figure 2 shows that the first stage fouling rates in the MBRs decreased linearly with increase in the shear intensities. Indeed, the high shear stress exerted on membrane fibers retard biofloc deposition and avoids sludge accumulation between fibers, particularly in the central region of the bundle. Wicaksana et al. (18) found that increased fiber movement induced by high air flow rate appeared to reduce the rate of biofloc deposition on the membrane surface and slow down the rise of TMP at fixed flux. The second stage fouling rates, as expected, were found to be significantly higher than that in the first stage. However, the second stage fouling rate in MBR₄₅₀ operation was found to be higher than that in MBR₃₀₀ which was indicative of optimum shear intensity in the MBR₃₀₀ as shown in Figure 2. At this point, it can be inferred that mixing intensity of certain extent is feasible to mitigate fouling, beyond which it becomes disadvantageous. The sludge characteristics of the MBRs were investigated to determine the influence of mixing intensity on deposited sludge cake properties.

Sludge filterability characteristics

Sludge filterability was characterized by the CST and the specific cake resistance (α). CST is a quantitative measure of the rate of water release from sludge and is indicative of the

filterability and dewaterability of sludge. In order to minimize the effect of suspended solids (SS) on CST, the CST values are normalized by dividing with MLSS concentration for each sample. In contrast, specific cake resistance (α) is a more authentic and reliable parameter for measuring the fouling potential or filterability of sludge cake. Figure 3 shows the averaged specific cake resistance (α) and the normalized CST (CST_N) for sludge samples from the MBRs. The filterability of sludge improved with increase in shear intensity up to 249 s^{-1} (MBR₃₀₀) in terms of both specific cake resistance (α) and CST_N . However, the filterability deteriorated for the MBR₄₅₀ sludge sample indicating that floc properties of MBR₃₀₀ exhibited lowest fouling potential. It can be inferred from the low CST_N and specific cake resistance (α) values of MBR₃₀₀ sludge that it was the appropriate hydrodynamic condition as well as the suitable sludge filterability characteristics that influenced the observed low fouling rates.

Particle size distribution (PSD)

Figure 4 (a) and (b) shows the PSD of the sludge in the MBRs within range $0.05\text{-}750\text{ }\mu\text{m}$ and $0.05\text{-}20\text{ }\mu\text{m}$, respectively. The median particle sizes were found to be 398, 379, 367 and $184\text{ }\mu\text{m}$ in the MBR₀, MBR₁₅₀, MBR₃₀₀ and MBR₄₅₀, respectively. Figure 4 (a) shows that the bio-particles became relatively smaller with increase in mixing rate from MBR₀ to MBR₃₀₀ with similar extent of distributions. However, MBR₄₅₀ exhibited significant reduction in particle sizes as well as scattered distribution. Moreover, Figure 4 (b) shows that the bio-particle distributions from MBR₀ to MBR₃₀₀ revealed similar trends within the range of $0.05\text{-}20\text{ }\mu\text{m}$ with exception of MBR₄₅₀ where increased percentage of particles greater than $10\text{ }\mu\text{m}$ suggested breakage of floc structure. The floc breakage into smaller particles under severe turbulent condition of MBR₄₅₀ could have induced the deterioration of the sludge filterability as depicted in Figure 3. However, the improved fouling potential of MBR₃₀₀ sludge could not be explained with the PSD results.

Bai and Leow (19) studied the effect of mechanical mixing intensity on membrane fouling in a cross-flow microfiltration system and observed that finer particles ($<50\text{ }\mu\text{m}$) caused severe membrane fouling. However, Sombatsompop et al. (20) found that the membrane fouling improved in an attached growth MBR system in the presence of smaller bio-particles ($17\text{-}33\text{ }\mu\text{m}$) as compared to larger particles ($65\text{-}226\text{ }\mu\text{m}$) in a suspended growth MBR system. Similarly, Lee et al. (6) found that in submerged MBR operation, the filtration performance enhanced with increase in air flow rate despite decrease in microbial floc size. Since activated sludge is a complex broth, it is difficult to explain the membrane fouling phenomenon explicitly on the basis of particle size.

Soluble and bound EPS

The soluble and bound EPS concentrations were determined by the addition of carbohydrate and protein concentrations measured in respective soluble and bound samples of the MBRs as shown in Figures 5 and 6, respectively. Figure 5 shows that the carbohydrate fraction of the soluble EPS was predominant among the MBR sludge. Moreover, soluble EPS concentrations in mechanically mixed MBRs were similar and slightly higher than that in the control system (MBR_0). Figure 6 shows that the bound EPS concentration in MBR_{450} was significantly higher as compared to that in the MBR_0 . The carbohydrate concentration in all the extracted samples was found to be similar but an increase was noticed in the protein levels of MBR_{450} . The high bound protein concentration in the MBR_{450} could be due to the bio-floc breakage releasing protein found at or outside the cell surface and in the intercellular space of microbial aggregate. The variation in the soluble and bound EPS concentrations, considered as major foulants, could not adversely influence the fouling mitigation achieved by the mechanical mixing condition in the MBR_{300} .

Discussion

The optimum shear intensity of 249 s^{-1} in the MBR₃₀₀ achieved low fouling rates in both stages. The first stage fouling was believed to be mitigated by the high shear intensity of mixed liquor turbulence inducing high fiber movement and slow deposition of biomass on the membrane fibers and between the fibers within the bundle. After cake formation in the second fouling stage, the fouling rate was improved by the high porosity and connectivity of deposited sludge cake depicted by the low specific cake resistance. The relationship between the specific cake resistances and the second stage fouling rates of the MBRs is shown in Figure 7. The linear curve shows a strong relationship between the specific cake resistances and the second stage fouling rates with an r-squared value of 0.99 suggesting the dependence of the second stage fouling rate on the specific cake resistance of a deposited cake layer. Thus, it can be postulated that specific cake resistance can be a reliable parameter to predict the extent of second stage membrane fouling rate in MBR filtration process.

CONCLUSION

The effect of sludge characteristics on membrane fouling were investigated in submerged MBRs operated at different mixing intensities. The membrane fouling behavior in terms of TMP variation with operating time was investigated. It was found that minimum fouling tendency was observed during MBR₃₀₀ operation with shear intensity (G) of 249 s^{-1} . Moreover, the CST_N and the specific cake resistance (α) of the sludge from MBR₃₀₀ were also found to be the lowest. Increase in G value beyond 249 s^{-1} was not able to retard fouling any further and the fouling potential deteriorated indicating mechanical mixing of 300 rpm as the optimum. A slight variation was observed in the particle size distribution from MBR₀ to MBR₃₀₀. On the contrary, the bioflocs broke into smaller particles in MBR₄₅₀ under extreme shear stress conditions. Based on these results, it can be postulated that improved filtration

performance of HF membranes can be achieved in submerged MBRs by physical modification of sludge properties and slow deposition of bioflocs on membrane surface induced by high mixing intensity. Moreover, increased fiber movement and homogeneous agitation by optimal shear intensity could avoid “dead zones” formation within the HF bundle.

Further investigation may be necessary into the microbial culture and activity and its impact on membrane fouling propensity to elaborate the present research findings of better MBR performance under appropriate mechanical mixing condition.

References

1. Stephenson, T.; Judd, S.; Jefferson, B.; Brindle, K. *Membrane bioreactors for wastewater treatment*; IWA Publishing: London, UK, 2000.
2. Zhang, J.; Chua, H.C.; Zhou, J.; Fane, A.G. Factors affecting the membrane performance in submerged membrane bioreactors. *J. Membr. Sci.* **2006**, *284*, 54.
3. Chang, I.-S.; Le-Clech, P.; Jefferson, B.; Judd, S. Membrane fouling in membrane bioreactors for wastewater treatment. *J. of Environ. Eng. (ASCE)*. **2002**, *128*, 1018.
4. Le-Clech, P.; Jefferson, B.; Chang, I. S.; Judd, S. Critical flux determination by the flux-step method in a submerged membrane bioreactor. *J. Membr. Sci.* **2003**, *227*, 81.
5. Germain, E.; Stephenson, T.; Pearce, P. Biomass characteristics and membrane aeration: Toward a better understanding of membrane fouling in submerged membrane bioreactors (MBRs). *Biotechnol. Bioeng.* **2005**, *90*, 316.
6. Lee, W.-N.; Kang, I.-J.; Lee, C.-H. Factors affecting filtration characteristics in membrane-coupled moving bed biofilm reactor. *Water Res.* **2006**, *40*, 1827.
7. Ji, L.; Zhou, J. Influence of aeration on microbial polymers and membrane fouling in submerged membrane bioreactors. *J. Membr. Sci.* **2006**, *276*, 168.

8. Yeo, A.P.S.; Law, A.W.K.; Fane, A.G. Factors affecting the performance of a submerged hollow fiber bundle. *J. Membr. Sci.* **2006**, *280*, 969.
9. Wang, X.-M.; Li, X.-Y.; Huang, X. Membrane fouling in a submerged membrane bioreactor (SMBR): Characterisation of the sludge cake and its high filtration resistance. *Sep. Purif. Technol.* **2007**, *52*, 439.
10. APHA. *Standard methods for the examination of water and wastewater*, 20th Ed.; American Public Health Association: Washington D.C., USA., 1998.
11. Frølund, B.; Palmgren, R.; Keiding, K.; Nielsen, P.H. Extraction of extracellular polymer from activated sludge using a cation exchange resin. *Water Res.* **1996**, *30*, 1749.
12. Dubois, M.; Gilles, K. A.; Hamilton, J. K.; Rebers, P. A.; Smith, F. Colorimetric method for determination for sugars and related substances. *Anal. Chem.* **1956**, *28*, 350.
13. Lowry, O.H., Rosebrough, N.R.; Farr, A.L.; Randall, R.J. Protein measurement with the folin phenol reagent. *J. Biol. Chem.* **1951**, *193*, 265.
14. Metcalf and Eddy. *Wastewater Engineering: Treatment and Reuse*, 4th Ed.; McGraw-Hill: New York, USA, 2003.
15. Cho, B.D.; Fane, A.G. Fouling transient in nominally sub-critical flux operation of a membrane bioreactor. *J. Membr. Sci.* **2002**, *209*, 391.

16. Le-Clech, P.; Chen, V.; Fane, A.G. Fouling in membrane bioreactors used in wastewater treatment. *J. Membr. Sci.* **2006**, *284*, 17.
17. Pollice, A.; Brookes, A.; Jefferson, B.; Judd, S. Sub-critical flux fouling in membrane bioreactors - a review of recent literature. *Desalination*. **2005**, *174*, 221.
18. Wicaksana, F.; Fane, A.G.; Chen, V. Fiber movement induced by bubbling using submerged hollow fiber membranes. *J. Membr. Sci.* **2006**, *271*, 186.
19. Bai, R.; Leow, H.F. Microfiltration of activated sludge wastewater-the effect of system operation parameters. *Sep. Purif. Technol.* **2002**, *29*, 189.
20. Sombatsompop, K.; Visvanathan, C.; Ben Aim, R. Evaluation of biofouling phenomenon in suspended and attached growth membrane bioreactor systems. *Desalination*. **2006**, *201*, 138.

List of Figures

Figure 1	TMP variation versus operating time during the MBRs operation
Figure 2	Fouling rates corresponding to shear intensities in the MBRs (G in MBR ₀ = 83 s ⁻¹ ; G in MBR ₁₅₀ = 117 s ⁻¹ ; G in MBR ₃₀₀ = 249 s ⁻¹ ; G in MBR ₄₅₀ = 439 s ⁻¹)
Figure 3	Specific cake resistance (α) and CST _N of the MBR sludge samples (G in MBR ₀ = 83 s ⁻¹ ; G in MBR ₁₅₀ = 117 s ⁻¹ ; G in MBR ₃₀₀ = 249 s ⁻¹ ; G in MBR ₄₅₀ = 439 s ⁻¹)
Figure 4 (a) & (b)	Particle size distribution of sludge suspensions in the MBRs (a) 0.05-750 μm and (b) 0.05-20 μm
Figure 5	Soluble EPS in mixed liquor of the MBRs
Figure 6	Bound EPS in mixed liquor of the MBRs
Figure 7	Relationship between Stage II fouling rate versus specific cake resistance (α) (G in MBR ₀ = 83 s ⁻¹ ; G in MBR ₁₅₀ = 117 s ⁻¹ ; G in MBR ₃₀₀ = 249 s ⁻¹ ; G in MBR ₄₅₀ = 439 s ⁻¹)

List of Tables

Table 1	Power requirement and velocity gradient expressions
Table 2	Shear intensity (G) in the MBRs

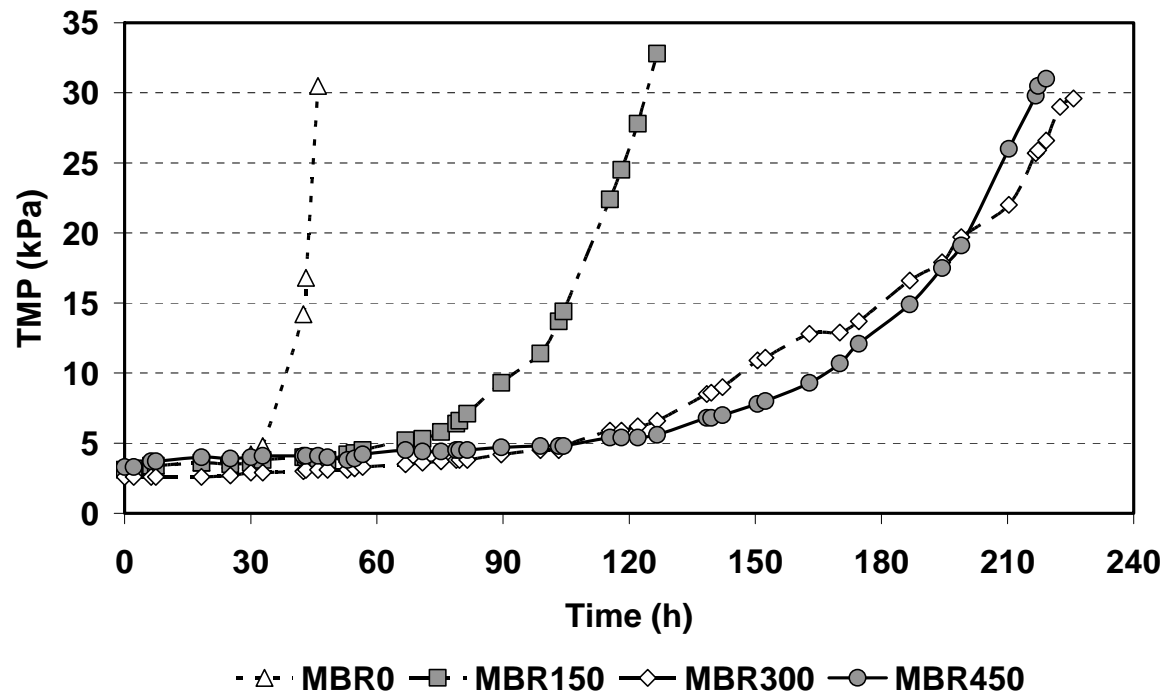


Figure 1. TMP variation versus operating time during the MBRs operation

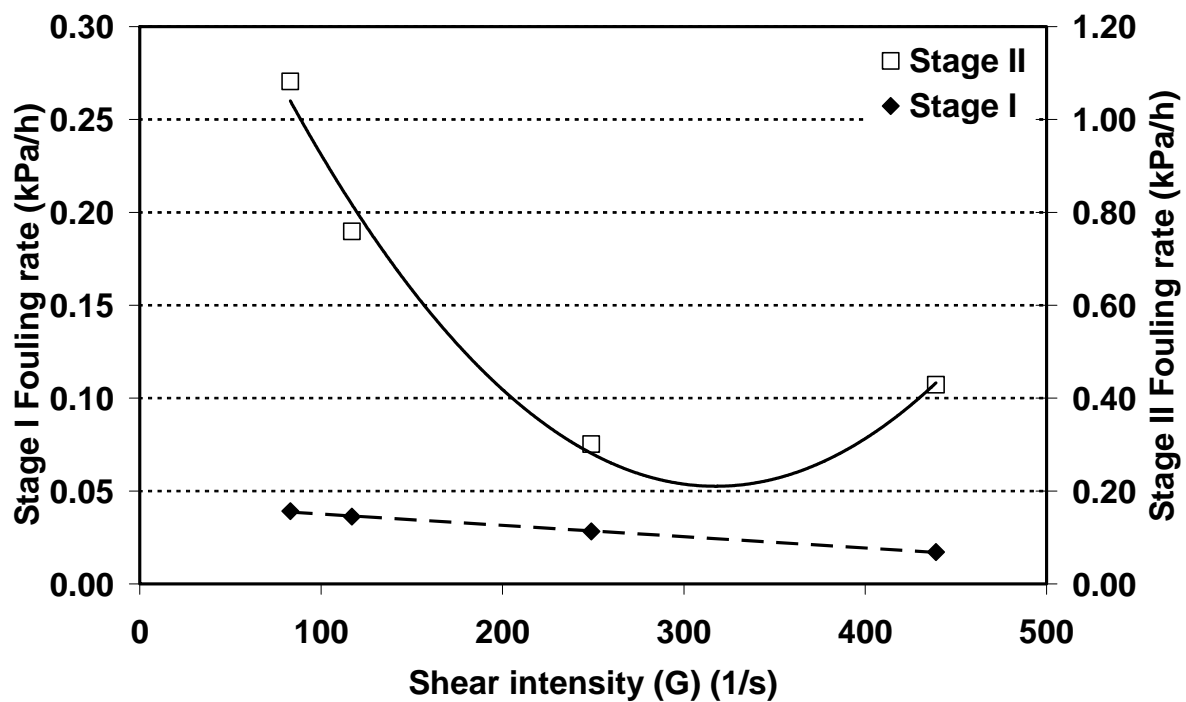


Figure 2. Fouling rates corresponding to shear intensities in the MBRs

(G in MBR₀ = 83 s⁻¹; G in MBR₁₅₀ = 117 s⁻¹; G in MBR₃₀₀ = 249 s⁻¹; G in MBR₄₅₀ = 439 s⁻¹)

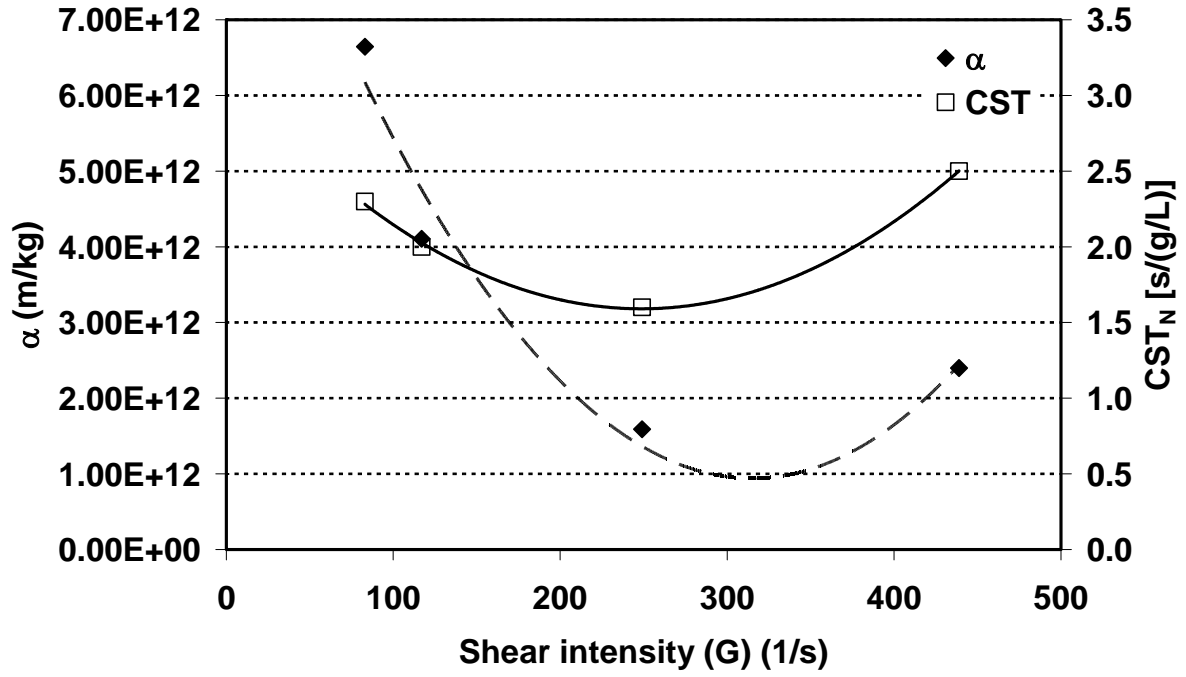


Figure 3. Specific cake resistance (α) and CST_N of the MBR sludge samples

(G in MBR₀ = 83 s⁻¹; G in MBR₁₅₀ = 117 s⁻¹; G in MBR₃₀₀ = 249 s⁻¹; G in MBR₄₅₀ = 439 s⁻¹)

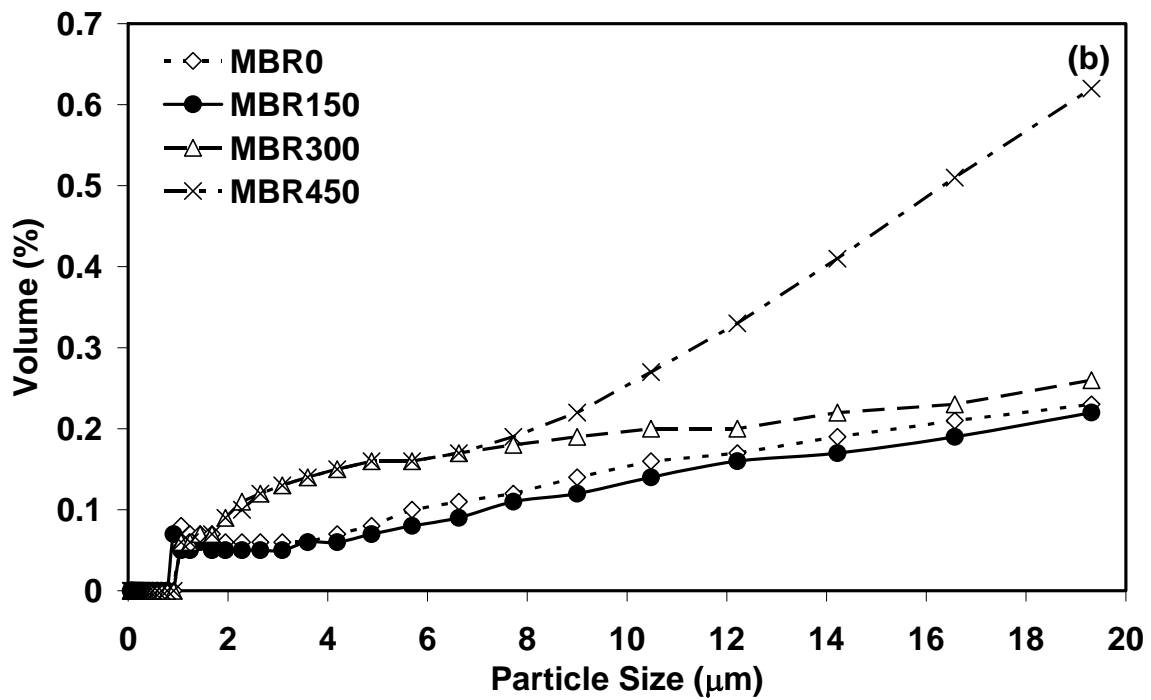
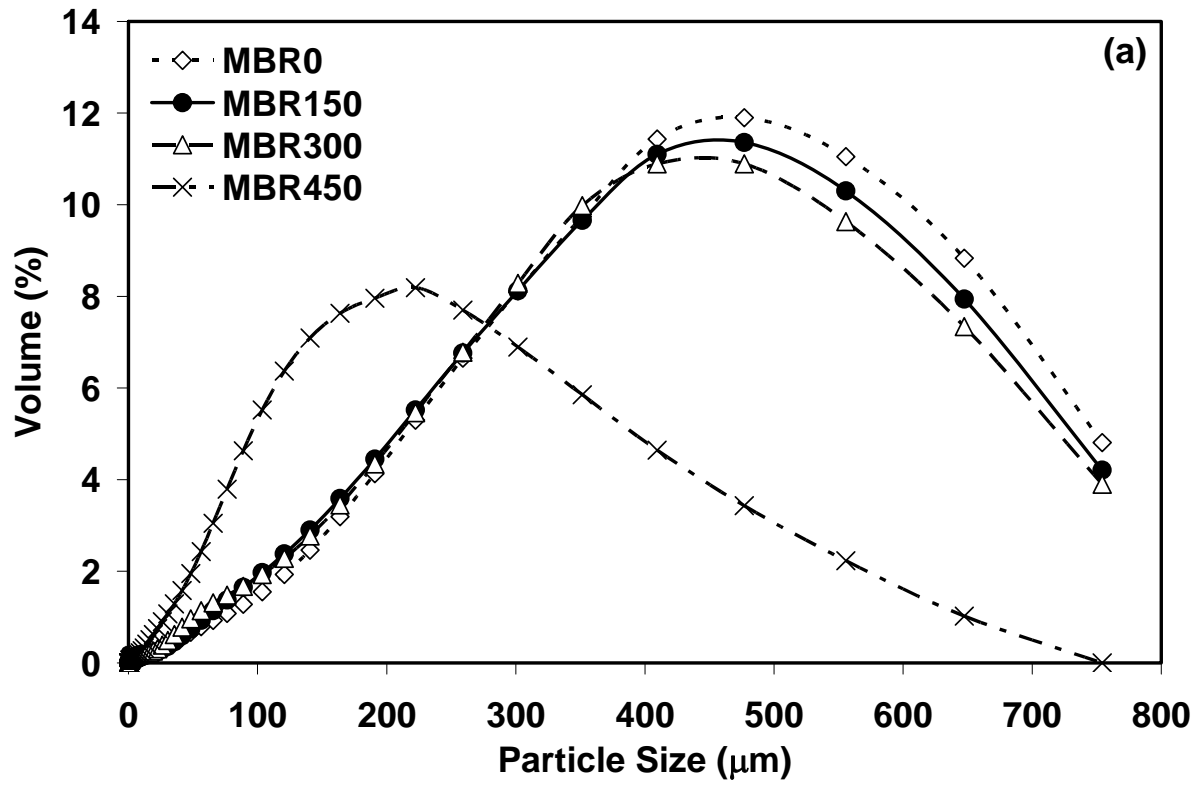


Figure 4. Particle size distribution of sludge suspensions in the MBRs (a) 0.05-750 μm and (b) 0.05-20 μm

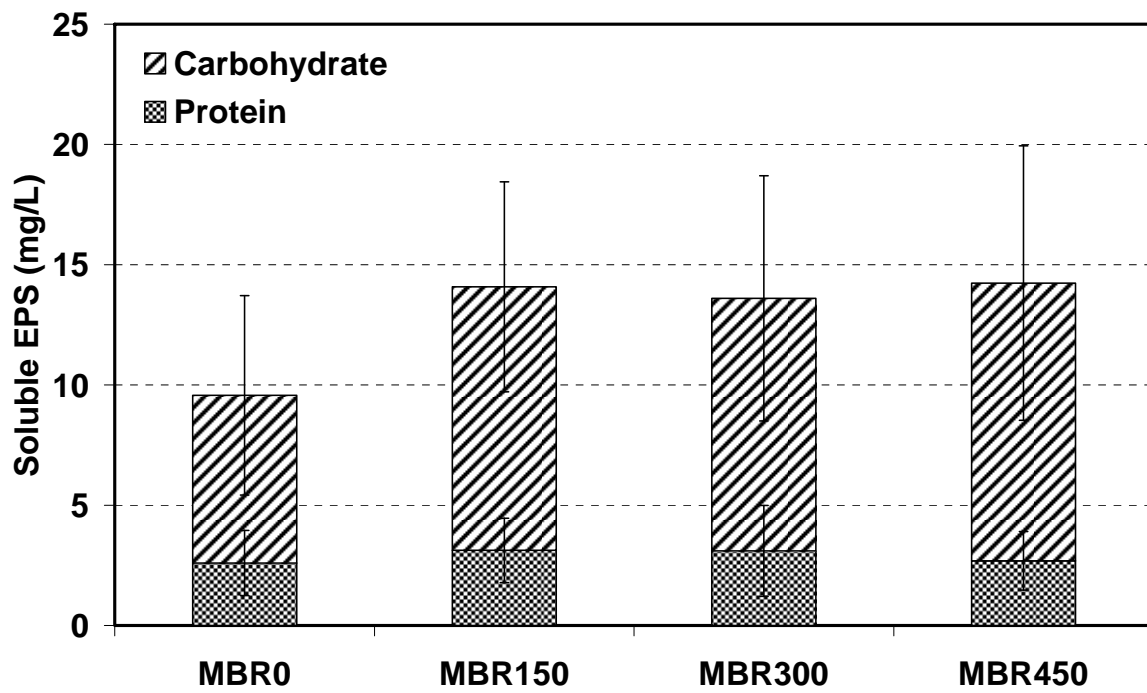


Figure 5. Soluble EPS in mixed liquor of the MBRs

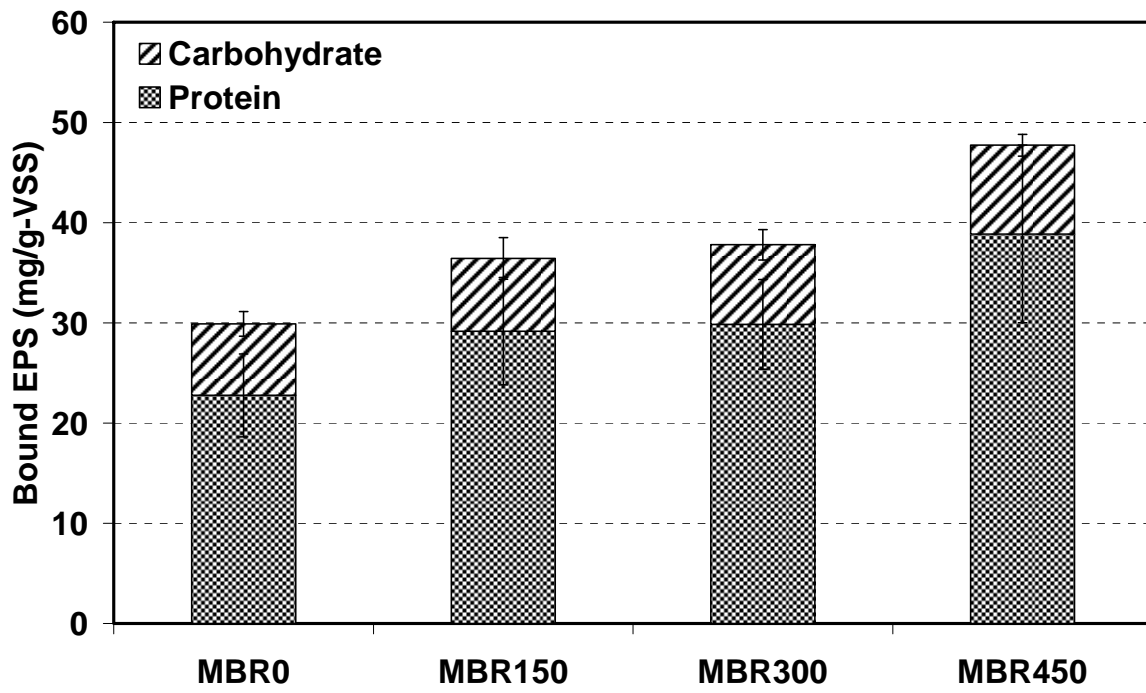


Figure 6. Bound EPS in mixed liquor of the MBRs

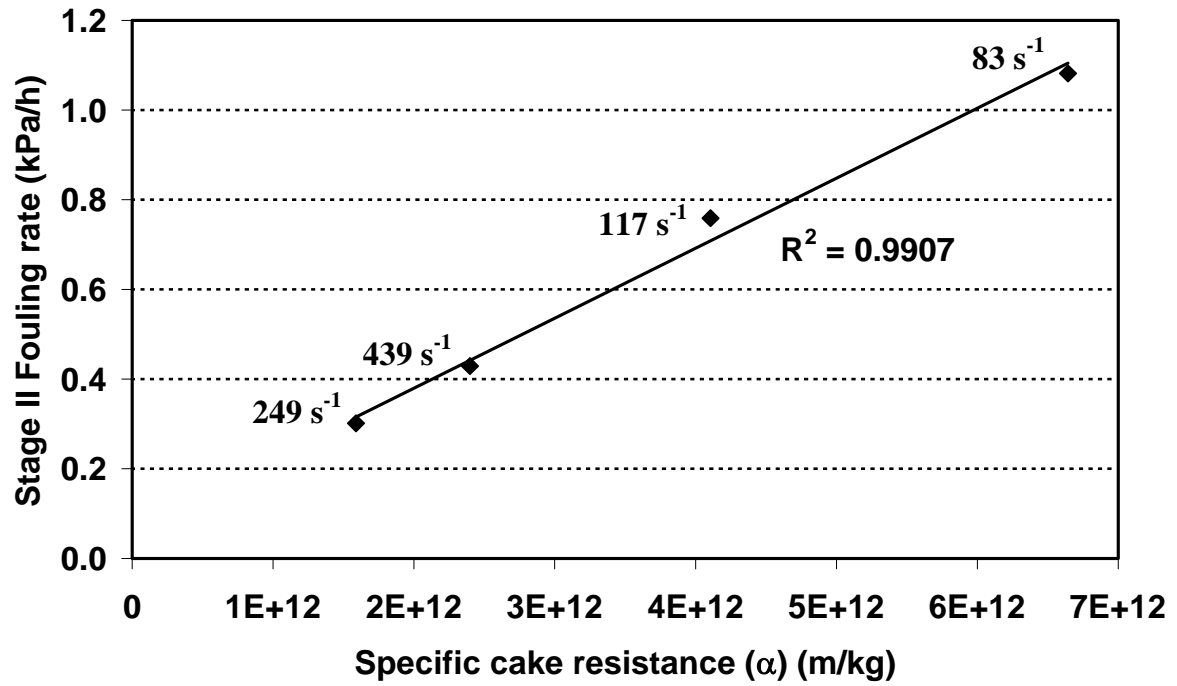


Figure 7. Relationship between Stage II fouling rate versus specific cake resistance (α)
 (G in MBR₀ = 83 s⁻¹; G in MBR₁₅₀ = 117 s⁻¹; G in MBR₃₀₀ = 249 s⁻¹; G in MBR₄₅₀ = 439 s⁻¹)

Table 1. Power requirement and velocity gradient expressions

Expression	Unit	Formula	Remarks
Mechanical power (P_m)	W	$P_m = N_p \rho n^3 D^5$ $N_R \geq 10,000$ $N_R = \frac{D^2 n \rho}{\mu}$	ρ = density of mixed liquor (1000 kg/m ³); n = mixing speed (rev/s) ; D = diameter of impeller (0.1 m); N_p = Power number for impeller ($N_p=1.1$); N_R = Reynolds number
Pneumatic power (P_p)	kW	$P_p = p_a V_a \ln \frac{p_c}{p_a}$	p_a = atmospheric pressure (kPa); V_a = air flow rate (m ³ /s); p_c = air pressure at the point of discharge (kPa)
Total power (P_T)	W	$P_T = P_m + P_p$	
Velocity gradient (G)	1/s	$G = \sqrt{\frac{P_T}{\mu V}}$	V = reactor volume (0.01 m ³); μ = dynamic viscosity (N-s/m ²)

Table 2. Shear intensity (G) in the MBRs

	Mechanical	Pneumatic	Reynolds	Power/				
	mixing	mixing	Number	P _m	P _p	Total	volume	
MBR	(rev/s)	(m ³ /h)	(N _R)	(W)	(W)	P (W)	(W/m ³)	G (1/s)
MBR ₀	0.0	0.3	0	0.00	0.17	0.17	17	83
MBR ₁₅₀	2.5	0.3	10,000	0.17	0.17	0.34	34	117
MBR ₃₀₀	5.0	0.3	20,000	1.38	0.17	1.55	155	249
MBR ₄₅₀	7.5	0.3	30,000	4.64	0.17	4.81	481	439

Influence of Mechanical Mixing Rates on Sludge Characteristics and Membrane Fouling in MBRs

**JAMAL KHAN Sher,
VISVANATHAN Chettiyappan,
JEGATHEESAN Veeriah,
BEN AIM Roger**

Presentation outline

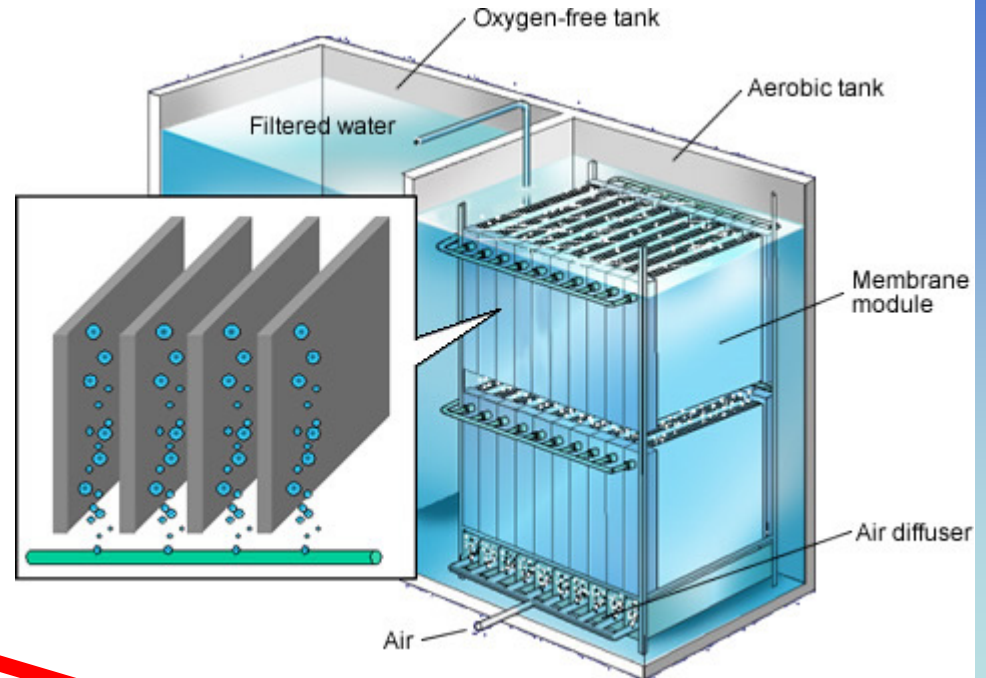
- Introduction
- Research objectives
- Methodology
- Results and Discussion
- Conclusion

Membrane bioreactor (MBR)

MBR = Combination of biological process by activated sludge + direct solid liquid separation by membrane filtration

Advantages:

1. High effluent quality
2. Good disinfection capability
3. High volumetric loading
4. Less sludge production
5. Small footprint & compactness

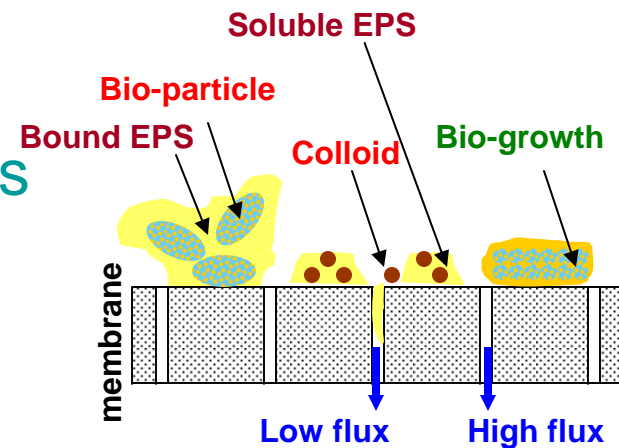


Membrane Fouling

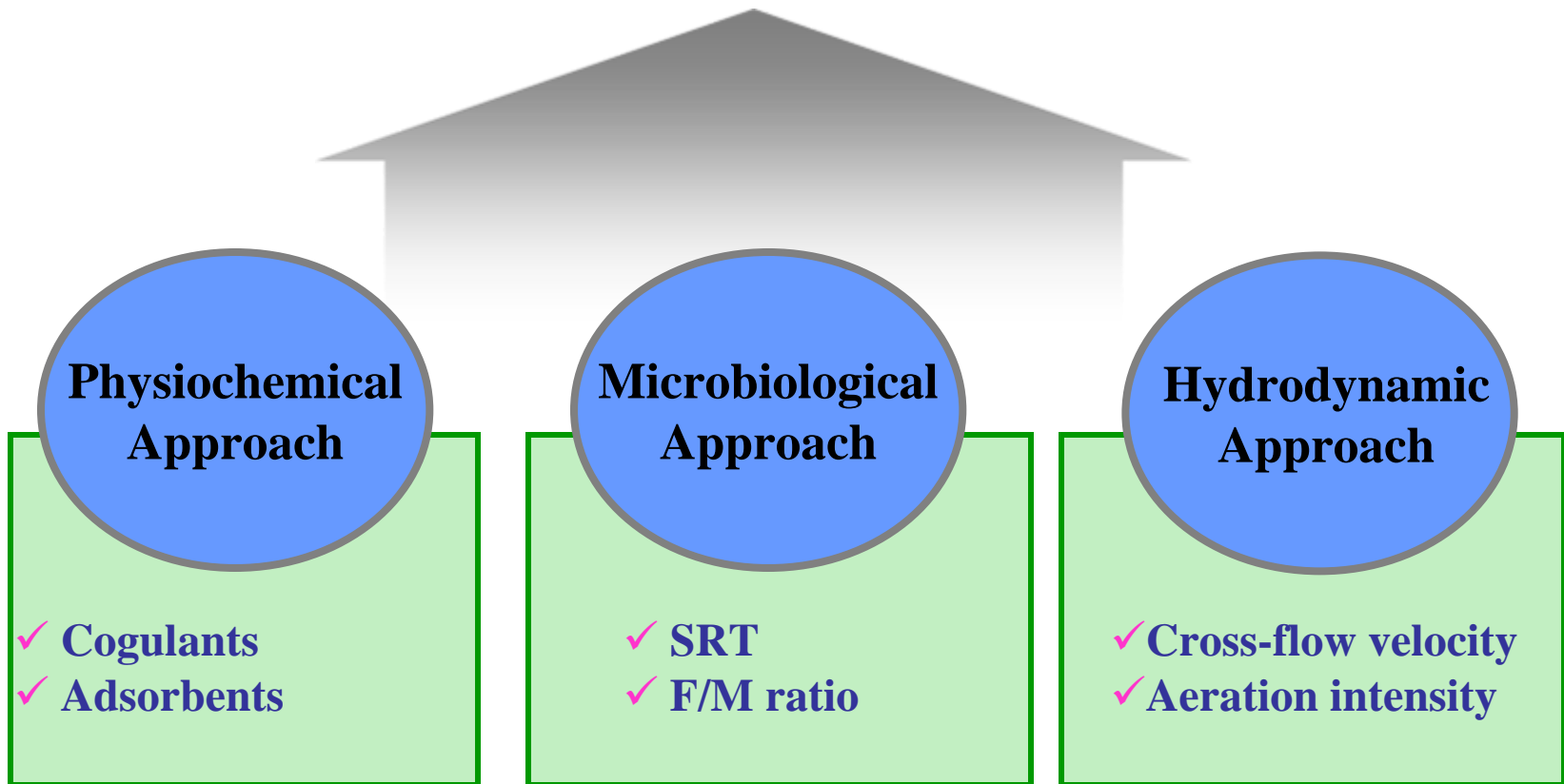
Accumulation of substances on membrane surface and/or within membrane pores resulting in deterioration of membrane performance

Major foulants:

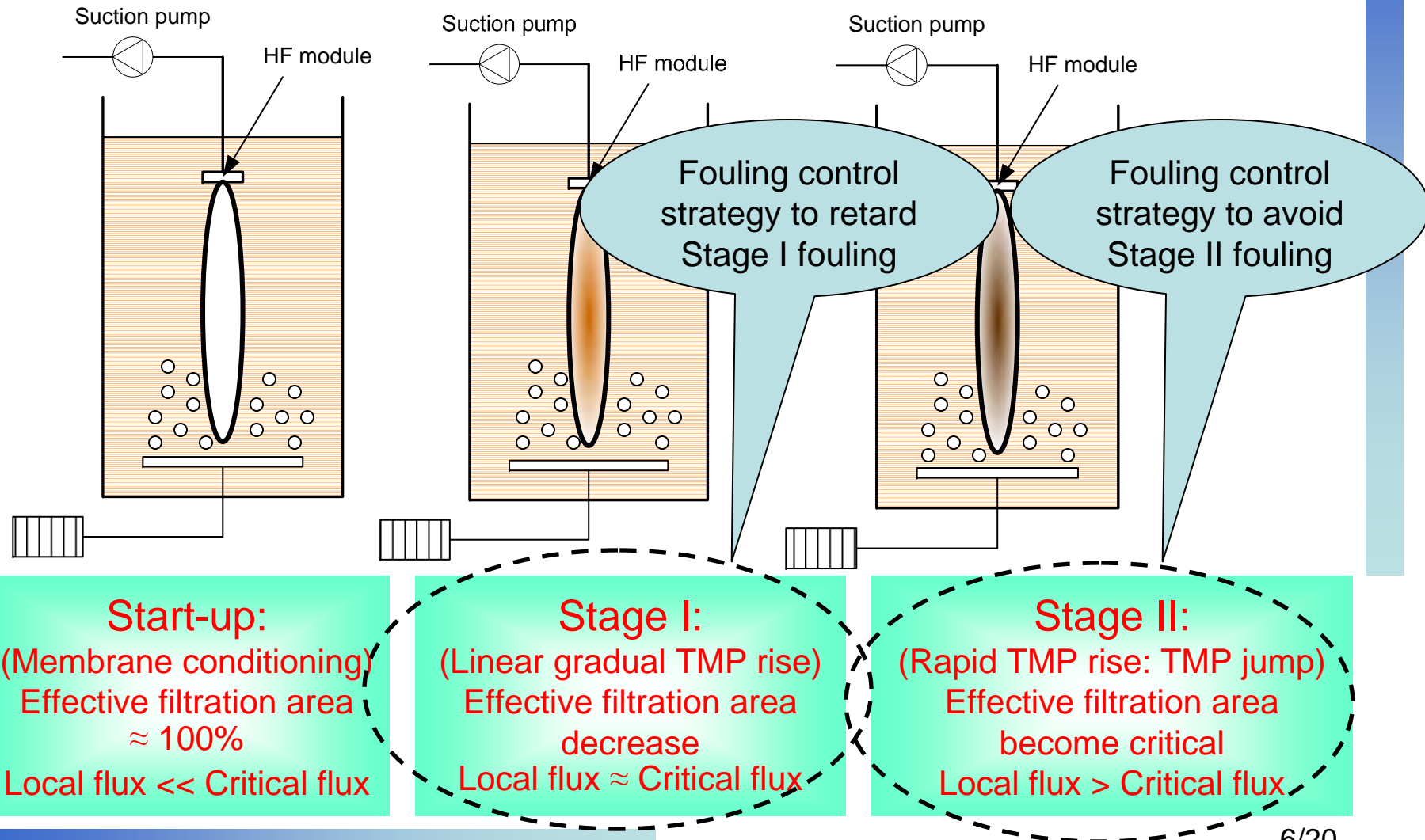
1. Suspended & colloidal particles
2. Soluble and bound EPS
3. Biological growth



Fouling control approaches



Biofouling in HF submerged MBR



Research background

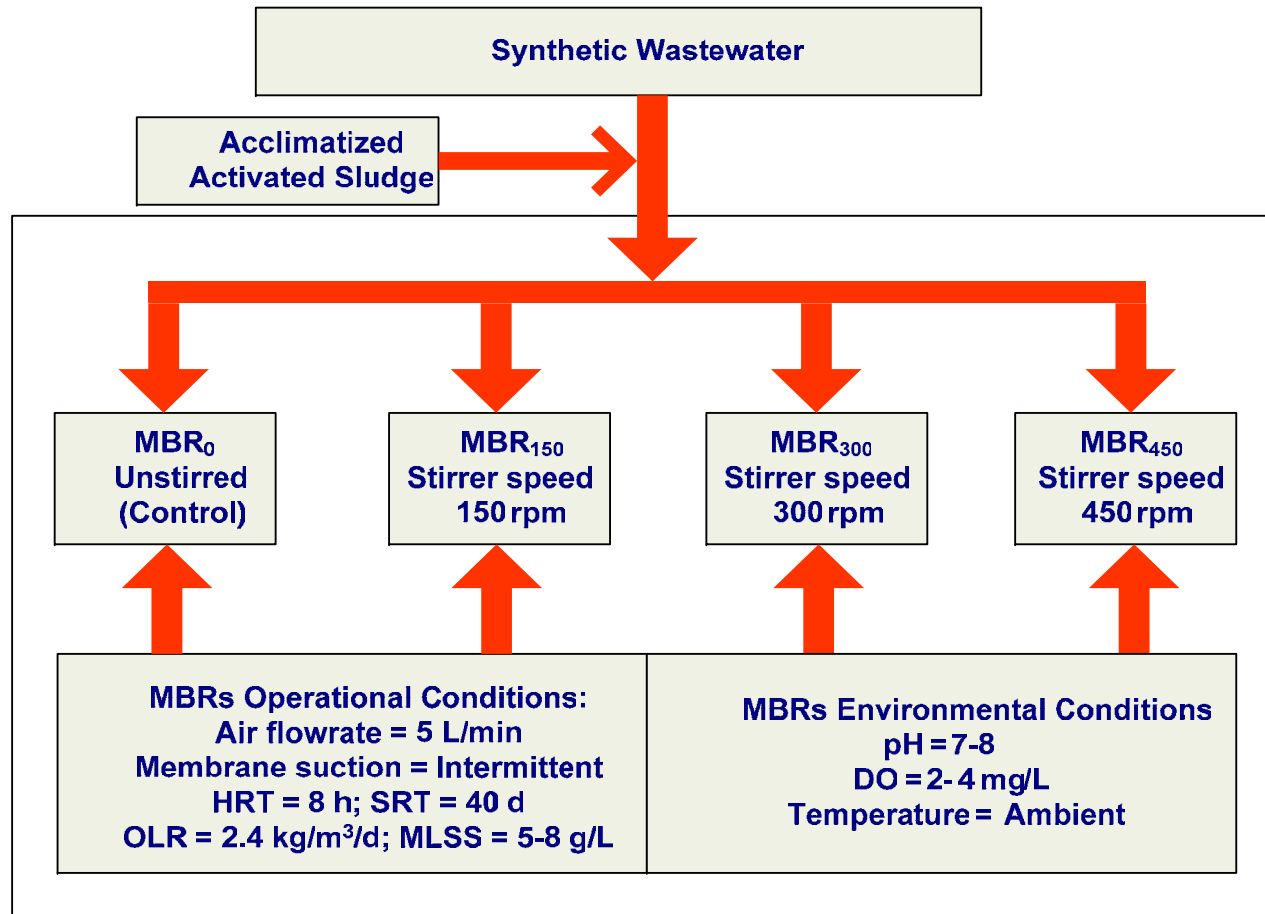
- High aeration intensity is preferred to hydrodynamically mitigate fouling in submerged MBR operation
- However, high aeration rates influence the biological conditions including
 - Growth rate
 - F/M ratio
 - Microbial community
- Biofloc deposit in low shear stress regions (vicinity of surrounded fibers) leading to local cake layer formation

High aeration rates can adversely effect the biological conditions as well as become ineffective with operational duration instigating the need to explore alternative hydrodynamic techniques

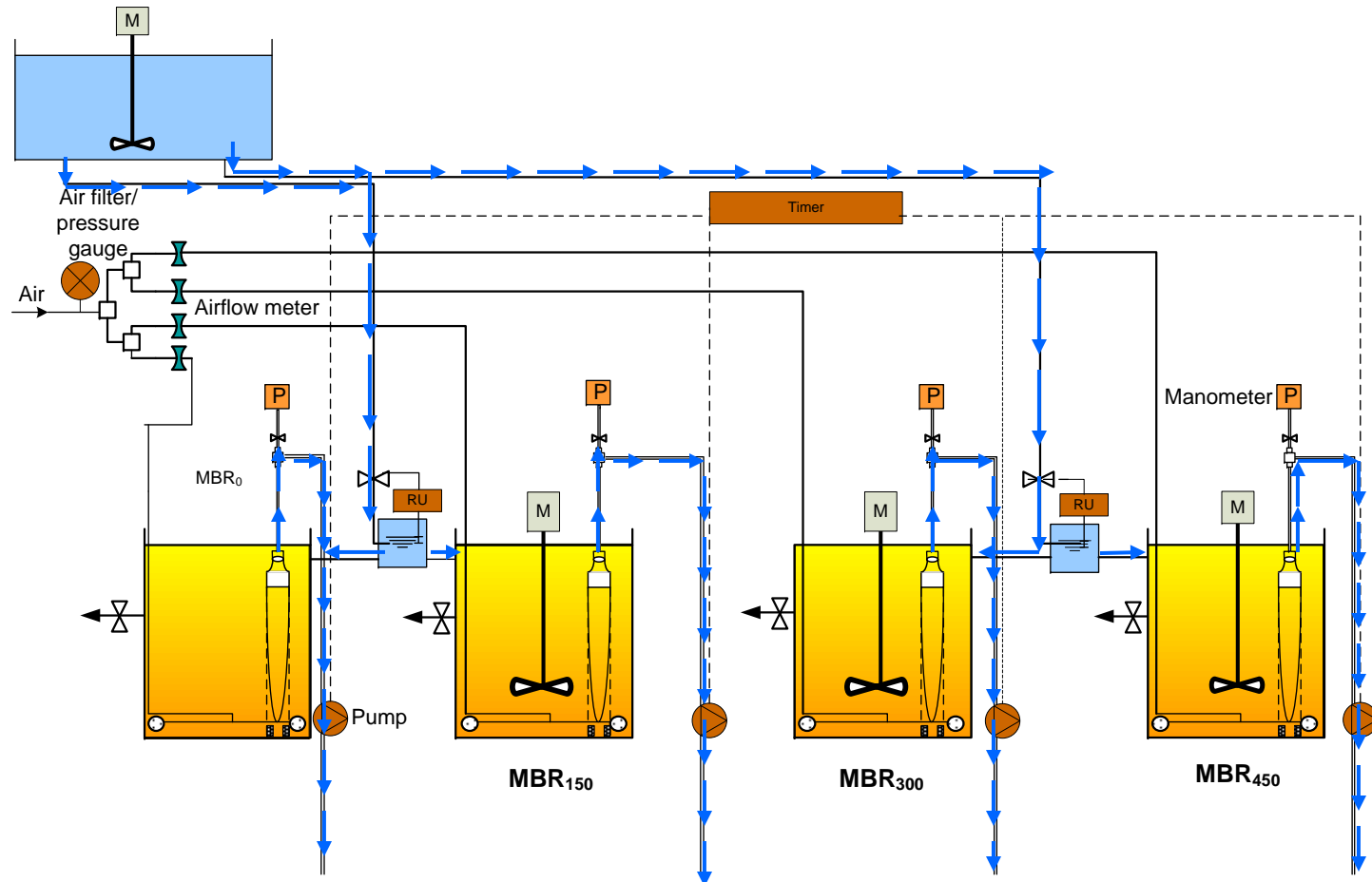
Research objectives

- Investigate mechanical mixing as an additional fouling control technique in MBR by modifying the hydrodynamic as well as biological environments
- Investigate sludge characteristics under variable mechanical mixing rates in MBRs
- Determine optimum mixing intensity in terms of improved filtration performance

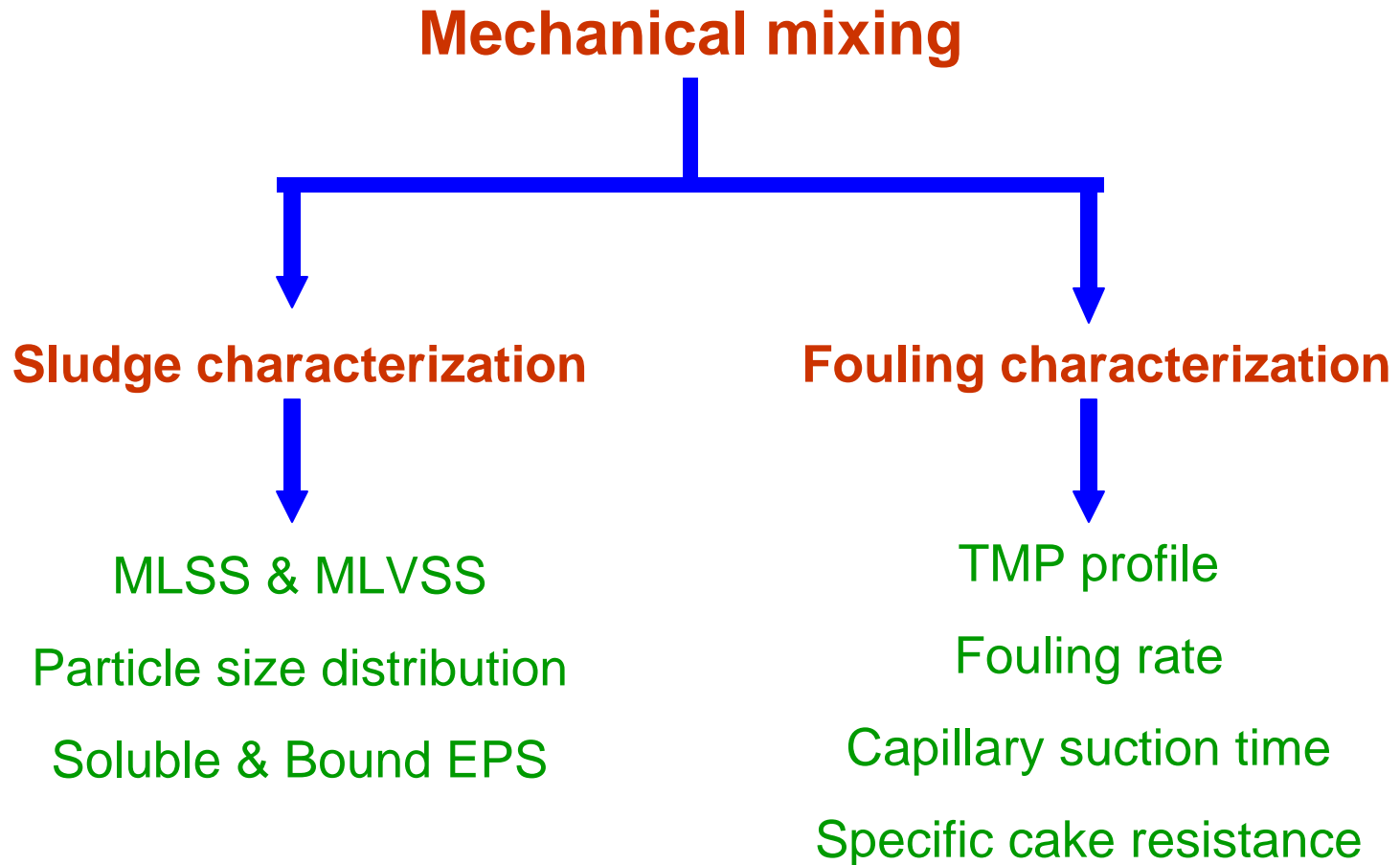
Operational conditions of MBRs



Experimental Setup



Research methodology

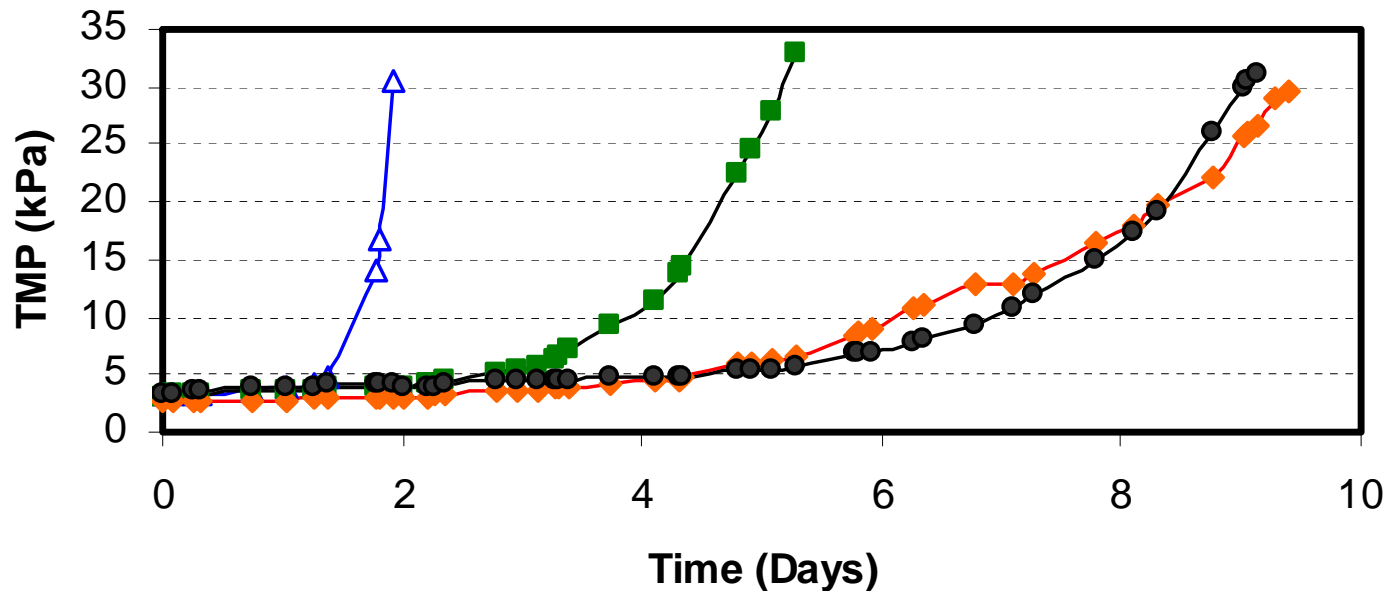


Mixing intensities in MBRs

Velocity gradient (G) in MBRs

MBR	Mechanical mixing (rev/s)	Pneumatic mixing (m ³ /h)	Reynolds Number (N _R)	Total power (W)	Velocity gradient (G) (1/s)
MBR ₀	0	0.3	0	0.17	83
MBR ₁₅₀	2.5	0.3	10,000	0.34	117
MBR ₃₀₀	5.0	0.3	20,000	1.55	249
MBR ₄₅₀	7.5	0.3	30,000	4.81	439

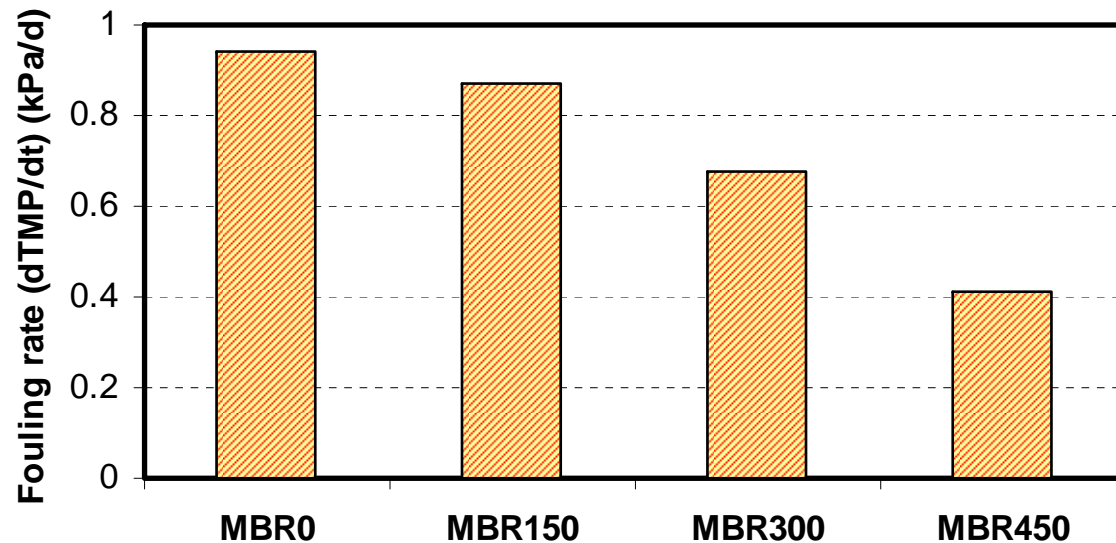
Membrane filtration performance



—△— MBR₀ —■— MBR₁₅₀ —◇— MBR₃₀₀ —●— MBR₄₅₀

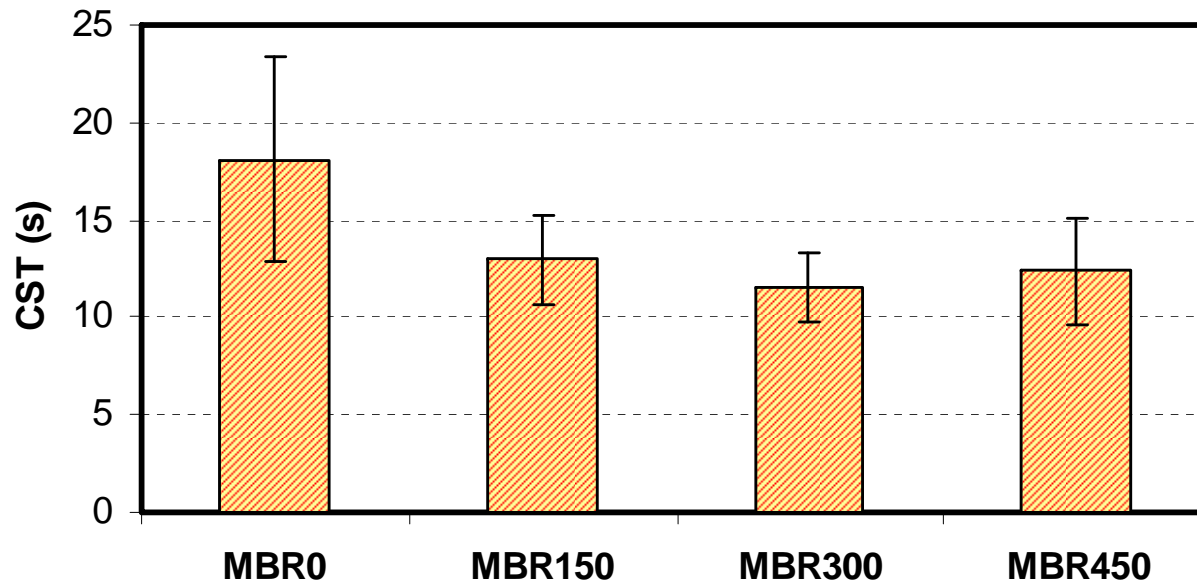
- MBR₀ fouled rapidly followed by MBR₁₅₀ and lastly by MBR₃₀₀ and MBR₄₅₀;
- MBR₃₀₀ exhibited least fouling due to appropriate mechanical mixing

Membrane fouling rates



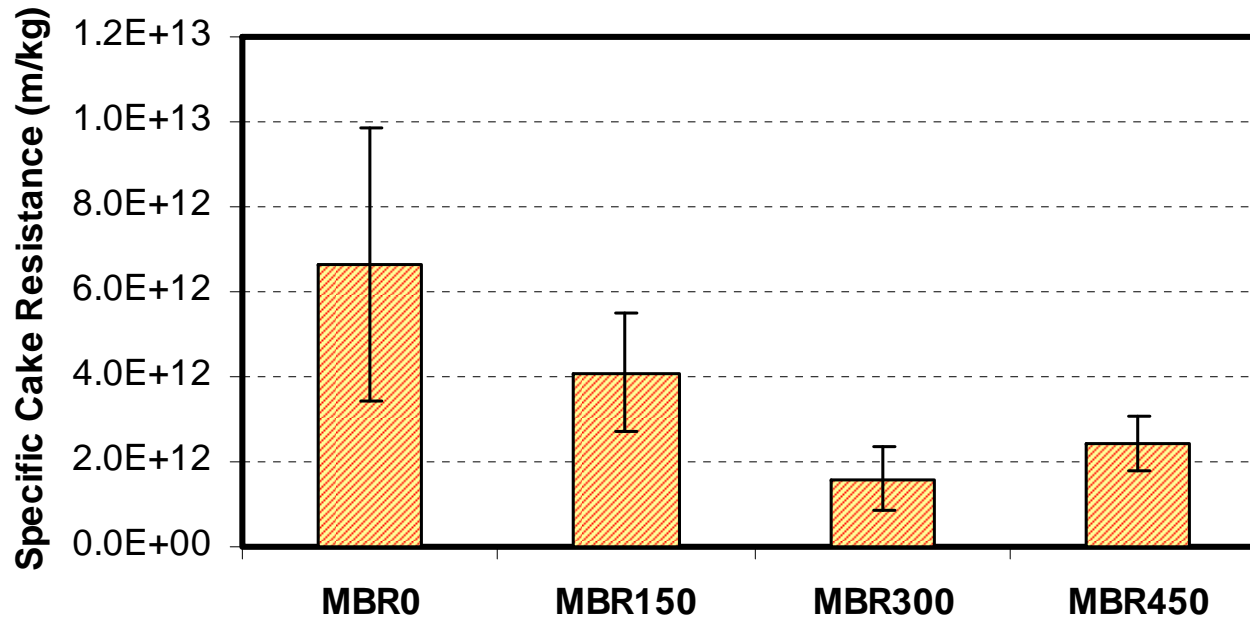
- Fouling rate characterized before rapid rise in TMP (7.0 kPa);
- Fouling rate decreased with increase in mixing intensity;
- MBR₃₀₀ and MBR₄₅₀ demonstrated low fouling rates

Capillary suction time (CST)



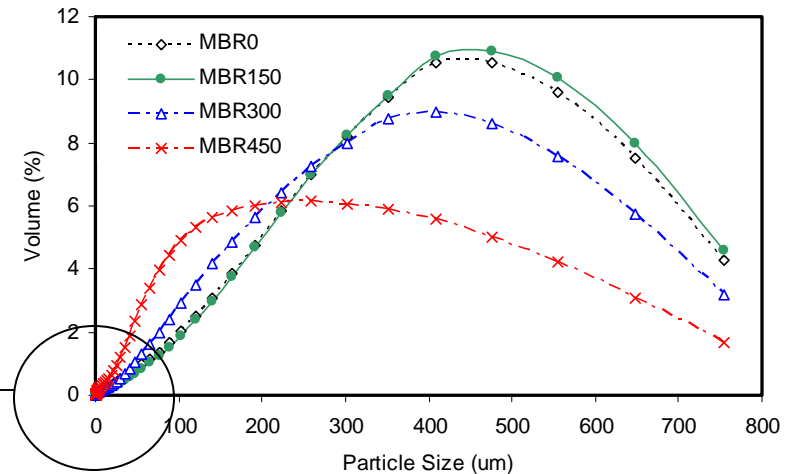
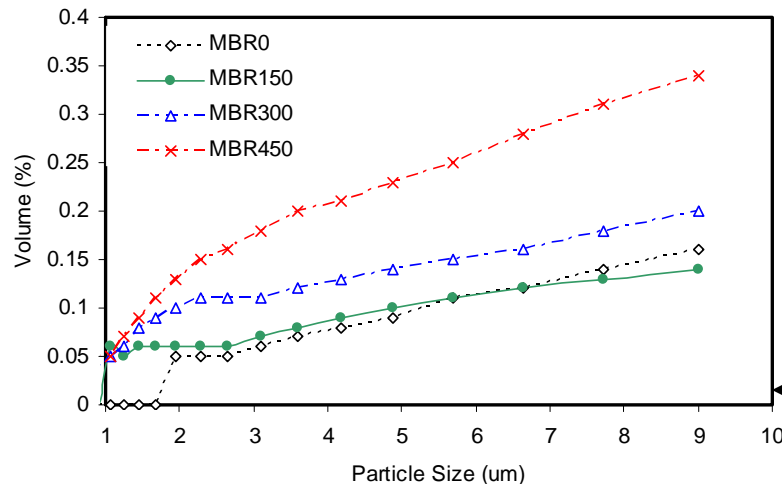
- CST indicated the filterability and dewaterability of the sludge;
- However, CST was relatively the same among the mechanically mixed MBRs;
- CST results could not establish relationship between sludge filterability and membrane filtration performance

Specific cake resistance ()



- MBR₃₀₀ sludge demonstrated the lowest fouling potential with 76% reduction in specific cake resistance as compared to MBR₀;
- Hydrodynamic shear stress on membrane fibers as well as high sludge filterability induced improved filtration performance in MBR₃₀₀

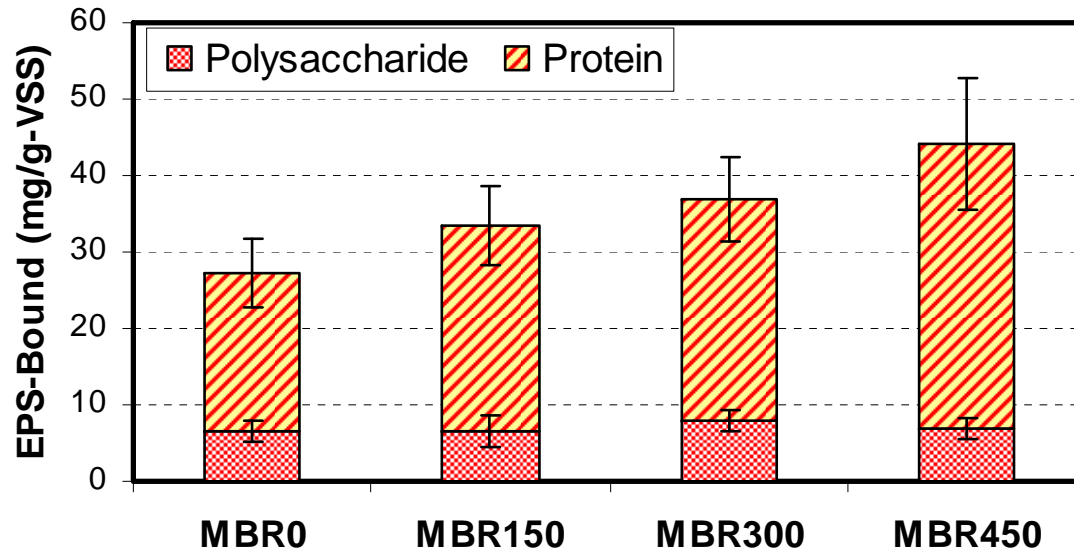
Particle size distribution



- Mean particle size reduced with increase in mixing intensity;
- Extreme turbulent condition (MBR_{450}) exhibited small particles and scattered distribution

MBR	Mean particle size (μm)
1 (Control)	375
2 (150 rpm)	385
3 (300 rpm)	333
4 (450 rpm)	244

Bound EPS



- Bound EPS concentrations increased in rapidly mixed MBRs as compared to that in conventional MBR;
- Protein content of bound EPS predominantly increased with increase in rapid mixing attributed to floc disintegration;
- EPS variation could not influence the membrane fouling rates

Conclusion

- Significant fouling reduction for optimal mechanical mixing condition of MBR₃₀₀ ;
- High filterability for MBR₃₀₀ sludge characterized by the specific cake resistance;
- Homogeneous shear stress distribution on membrane fibers and small bio-particles induced improved filtration performance under optimal mixing condition of MBR₃₀₀

**Thank you
for your attention**